環境システム研究 Vol. 24 1996年10月

Modeling of algae blooming and macrophyte development in Lake Veluwe

Truong Van Bon* and Takashi Asaeda**

Abstract; A model of three species of phytoplankton (diatom, green and blue green algae) coupled with macrophyte and sedimentary model has been developed for shallow lake. Phosphorus and nitrogen return fluxes from the bottom sediment to overlying water are simulated by model for the sediment -water exchange of nutrients. The development of potamogeton pectinatus L. is simulated by macrophyte model. The verification of the model for phytoplankton, phosphorus, ammonium, nitrate return fluxes was conducted using observed data in Lake Veluwe in 1986. The fitness between observed and its computed values suggest that the model is capable of reproducing the real picture of phytoplankton and Potamogeton P.L. development in Lake Veluwe for short and long term time period. One of the rationale conclusion from the model for Lake Veluwe is that the declining of phytoplankton concentration followed the water quality is improved, as consequences light can be penetrated into deeper water column and macrophyte can be grown. All model equations and the application of the model as tool for analyzing and describing interactions between the biological processes, nutrients recycling, macrophyte developing in Lake Veluwe is subject of this study.

1. Introduction

Lake ecosystem experiences large fluctuation in submerged macrophyte biomass (Carpenter and Lodge, 1986). Submerged macrophyte affects on ecosystem through physical, chemical, biological processes and vice versa macrophyte development is determined by environmental conditions. In shallow eutrophic lakes where algae concentration is abundant, light penetration is inhibited and, thus becomes determinant condition for macrophyte development. Macrophyte biomass will result in changes in production of the ecosystem (Carpenter and Lodge, 1986). Heavy crops of macrophytes inhibit the production of phytoplankton by shading effects. When macrophytes are dominant, phytoplankton production is reduced by as much as two orders of magnitude for periods, and the macrophyte crop may be reduced in years with persistent large phytoplankton crops (Mitchell, 1989). In recent years, the restoration of highly turbid, eutrophic waters is highly demanded. Since the understand of the interactions between ecological components is essential, mathematical model is available as a tool for solving problems encountered in the management (Hootsmans, 1994). This study is devoted to develop a numerical model linking submerged macrophytes Potamogeton pectinatus L. under the development of three species of phytoplankton (diatom, green and blue green algae) and the exchanges of nutrients between the sediment and overlying water.

2. Model description

Ecological model consists three submodels that describe phytoplankton development, nutrient dynamics in the sediment and in the overlying water, and subnerged macrophyte growth in a shallow lake. These models are summarized as follows.

2.1 Phytoplankton model

Three species of algae (diatom, green, blue green) are considered in the model. Chlorophyll-a is considered as an indicator of phytoplankton biomass both for total phytoplankton concentration or for single species. Phytoplankton growth follows the limitation by the light intensity according to Steele function (1962), by the internal phosphorus, and the internal nutrient to functions proposed by Hamilton and Schladow (1994a) and by the external silica concentration for diatom following a Michaelis-Menten equation, and the environmental temperature function by Eppley (1972). Phytoplankton biomass is resulted by four processes: growth, respiration, mortality and zooplankton grazing. The model equations and the explanations of the variables are listed in table 1. There are 14 state variables in this submodel.

Graduate Student, Dept. of Civil and Env. Eng., Saitama University, 338 Japan

Associate Professor, Dept. of Civil and Env. Eng., Saitama University, 338 Japan

Table 1. Model equations of phytoplankton and nutrient concentrations in the water column.

$$\frac{E\Delta h_i}{a} - G_{\text{max}_i} q^{T-20}Ck_i, \text{ male}_i^T(J_i), f(P_i), f(N_i), f(S_i)_i^T \} k_{xxx_i} q^{T-20}Ck_i^L - k_{mor_i} q^{T-20}Ck_i^L - k_{mor_i}$$

2.2 Sediment -water exchange model for nutrients

A dynamic model for prediction of nutrient fluxes across the sediment - water interface has developed in order to investigate the nutrient fluxes release from the sediment. In a natural lake, it is observed that the activity of benthic fauna plays an important role in the determination of vertical characteristics of the active bottom layer. The main transport processes in sediment are advection due to infiltration, seepage or incorporation, diffusion, bioturbation and at the interface, sedimentation and resuspension (Smits et al., 1993; Chapelle, 1994). In the model, the sediment layer under consideration is divided into four sublayers (table 2). The aerobic layer, the denitrifying layer, the upper reduced layer and the lower reduced layer. A thin stable boundary layer is supposed to exist over the sediment - water interface. The equations and the explanations of the state variables are summarized in table 3. A set of 12 state variables of the submodel needs to be solved.

Table 2. Schematization of processes in the lake -bottom sediment and overlying water

Layer	Detritus	Nitrate	Ammonium	Dissolved Oxygen	Phosphate
ZO Water column	Settling IncorporationMi neralization Degradation Humification	Denitrification Nitrification	Ammonification Nitrification Adsorption Dispersion	Photosynthesis Oxydation BOD Reacration	Mineralization Sorption Dispersion
Newly deposited detritus	Degradation Incorporation		Ammonification	aoo.	Release
zl aerobic layer		Dispersion Nitrification Flux to detrinifying layer	Seepage Degradation Nitrification Dispersion Scopage	Degradation	Adsorption Seepage Mineralization Bioturbation
22 denitrifying layer			Degradation	Nitrification	Dispersion
z3 upper reduced layer	Bioturbation	Denitrification	Scopage	Deoxygenution	Secpage Burial Mineralization Biotubation Diapersion
z4 lower reduced layer	Burial Degradation	Denitrification	Degradation	Deoxygenation	Scepage Burial Mineralization
Inactive bottom layer	Inactive	Inactive	Scopage	Inactive	Seepage Burial

Table 3. Model equations of sediment - water exchange of nutrient.

```
\frac{dDet_o}{-q.a.M.Chl_a-u.Det_o-s.Det_o}; z_1 = \sqrt{2.\alpha_1.D_{oxy}.k_0.0/OC};
                                                                                                                                                    Fa_3 = (a_1 + a_4)D_{oran} \frac{NH_4 - NH_2}{z_1 + z_3 + z_4 - z_1}; PO_{ods} = f_{ads}PO; PO_{ora} = f_{ora}PO
OC = OC_b / z_1 + ac.k_{deg1} \cdot Det_1 + p_1 \cdot an.k_{nit1} \cdot NH_1 + COD / z_1
COD = (1 - fra).ac. [k_{deg1}. Det_1(z_1 + z_3) + k_{deg4}. Det_4.z_4]
                                                                                                                                                                   Fp_{boun} + Fp_{sed} - Fp_{wol} + Fp_2 + Fp_{seepo} - Fp_{seep2} + Fp_{bio2} - Fp_{ber2} \cdot PO_1
z_2 = 2.(NO_1 - NO_c)/NO_1 \cdot \sqrt{D_{nur}/k_{den}}; OC_b = ac.k_{degb} \cdot Det_b
\frac{dDel_b}{dt} = Fd_{Nt} - Fd_{co} - k_{deqb} \cdot Del_b \cdot \frac{dDel_1}{dt} = \frac{Fd_{co} - F_{burg} \cdot Del_1}{z_1 + z_2 + z_3} - k_{deqf} \cdot Del_1
                                                                                                                                                    +ap.kdeg1.PO1
                                                                                                                                                    \frac{dPO_3}{dPO_3} = \frac{-Fp_2 + Fp_3 + Fp_{seep2} - Fp_{seep3} - Fp_{bio2} + Fp_{bio3} + Fp_{ber2} \cdot PO_1 - Fp_{ber3}}{-Fp_{bio2} + Fp_{bio3} + Fp_{ber2} \cdot PO_1 - Fp_{ber3}}
\frac{dDet_4}{dDet_4} = \frac{F_{ber3} \cdot Det_1 - Fd_{bio3}}{Fd_{co} - rc. Det_6} - k_{deg4} \cdot Det_4; Fd_{ni} - sc. Det_6; Fd_{co} - rc. Det_6
                                                                                                                                                     +ap.k<sub>deg1</sub>.PO<sub>3</sub>
Fd_{be3} = \frac{2D_b}{z_1 + z_2 + z_3 + z_4} \frac{Det_4}{1 - \alpha_4} \frac{Det_1}{1 - \alpha_1}
                      2 D<sub>b</sub>
\frac{dNO_{1}}{dNO_{2}} = \frac{-Fn_{was} + Fn_{1}}{-Fn_{was} + K_{mil} \cdot NH_{1} - K_{dent} \cdot NO_{1}}
                  \alpha_1.z_1
               \frac{-Fn_1 + Fn_3}{\alpha_1 \cdot (z_2 + z_3)} + k_{nit2} \cdot NH_2 - k_{den2} \cdot NO_2
dNO<sub>2</sub>
  đt
\frac{dNO_4}{dl} = \frac{-Fn_3}{\alpha_4 \cdot z_4} + k_{nll4} \cdot NH_4 - k_{den4} \cdot NO_4; \ Fn_3 = 2 \cdot (\alpha_1 + \alpha_4) \cdot D_{nlm} \cdot \frac{NO_4 - NO_2}{z_2 + z_3 + z_4}
                                                                                                                                                    F_{P2} = 2.D_{pq} \cdot \frac{f_{dis3}PO_3 - f_{dis1}PO_1}{k}; \ k = k_{20} \cdot k'^{-20}
                                                                                                                                                                                  z_1+z_2+z_3+z_o
Fn_{\omega \sigma} = 2.\alpha_1.D_{nw} \frac{NO_1 - NO_{\omega \sigma}}{NO_2 - NO_1}; Fn_1 = 2.\alpha_1.D_{nw} \frac{NO_2 - NO_1}{NO_2 - NO_2}
                                                                                                                                                                                         Sdis4.PO4 Sdis3.PO3
                                                                                                                                                    Fp_3 = (\alpha_1 + \alpha_4) D_{po}. \frac{\alpha_4}{z_3 + z_4}
           = \frac{Fa_b - Fa_{ww} + Fa_1 + Fa_{nego} - Fa_{negl}}{Fa_b - Fa_{ww} + Fa_1 + Fa_{nego} - Fa_{negl}} + \frac{aa.k_{degl} \cdot Det_1}{A_{obs} \cdot Det_2} - k_{ww} \cdot NH_1
                p_1.d_1 - Fa_1 + Fa_3 + Fa_{seep1} - Fa_{seep3} +
  dt
 dNH_2
                                                                            aa.kdegl.Detl
                   \alpha_1.(z_2+z_3)
   dí
               -Fa3 - Fangs - Fangs - Fangs + aa.k deg 4. Del 4
dNH₄
                          \frac{\alpha_{4}.z_{4}}{\frac{NH_{1}-NH_{1}}{z_{1}+z_{2}}}; PO_{pu}-f_{pu}PO; PO_{dig}-f_{dig}PO/\alpha
  dt
```

 $\frac{(z_1 + z_2 + z_3)(1 - \alpha_1)}{(\zeta_{pre4} + \zeta_{ads4})PO_4} \left(\zeta_{pre3} + \zeta_{ads3} \right)$ Chi_a chloropyll-a concentration in overlying water (mg m⁻³); Det, detritus concentration in overlying water (g C m⁻³); Det, detritus concentration in boundary layer (g C m⁻²); Det detritus concentration in the sediment (g C m⁻³); NO nitrate concentration in the sediment (g C m³); NH ammonium concentration in the sediment (g C m³); PO_{st}, adsorbed phosphate concentration in the sediment (g P m³); PO_{st}, dissolved phosphate concentration in the sediment (g P m³); PO_{pr}-precipitated phosphate concentration in the sediment (g P m³); PO_{pr}-organic phosphate concentration in the sediment (g P m⁻³); Fa_{bon} ammonification flux from the boundary layer (g N m⁻²d⁻¹); Fa_{bon} bioturbation flux of detritus (g C m²d¹); FPss bioturbation flux of phosphate between layers (g P m²d¹); COD chemical oxygen demand (g O2 m²d¹); Fa. dispersive ammonium flux between layers (g N m²d⁻¹); Fo_w dispersive ammonium return flux to overlying water (g N m²d⁻¹); Fo_w-dispersive nitrate return flux to overlying water (g N m²d⁻¹); Fp, dispersive phosphate flux between layers (g P m²d⁻¹); Fp_w dispersive phosphate return flux to overlying water (g P m²d⁻¹); PO_m dissolved phosphate concentration (g P m²d⁻¹); Fd_m flux of detritus incorporated in the upper layer (g C m²d⁻¹); Fd_m flux of detritus settled from overlying water (g C m²d⁻¹); Fd_m flux of detritus settled from overlying water (g C m²d⁻¹); Fd_m flux of detritus settled from overlying water (g C m²d⁻¹); Fd_m flux of detritus settled from overlying water (g C m²d⁻¹); Fd_m flux of detritus settled from overlying water (g C m²d⁻¹); Fd_m flux of detritus settled from overlying water (g C m²d⁻¹); Fd_m flux of detritus settled from overlying water (g C m²d⁻¹); Fd_m flux of detritus settled from overlying water (g C m²d⁻¹); Fd_m flux of detritus settled from overlying water (g C m²d⁻¹); Fd_m flux of detritus settled from overlying water (g C m²d⁻¹); Fd_m flux of detritus settled from overlying water (g C m²d⁻¹); Fd_m flux of detritus settled from overlying water (g C m²d⁻¹); Fd_m flux of detritus settled from overlying water (g C m²d⁻¹); Fd_m flux of detritus settled from overlying water (g C m²d⁻¹); Fd_m flux of detritus settled from overlying water (g C m²d⁻¹); Fd_m flux of detritus settled from overlying water (g C m²d⁻¹); Fd_m flux of detritus settled from overlying water (g C m²d⁻¹); Fd_m flux of detritus settled from overlying water (g C m²d⁻¹); Fd_m flux of detritus incorporation (g C m²d⁻¹); Fd_m flux of detritus settled from overlying water (g C m²d⁻¹); Fd_m flux of detritus settled from overlying water (g C m²d⁻¹); Fd_m flux of detritus incorporation (g C m²d⁻¹); Fd_m flux of detritus incorpora rate of detritus in the sediment (d-1); so sedimentation rate of detritus (m d-1); t time; ws specific weight of the sediments (kg m-3); op stoichiometric constant for phosphate in detritus (g P/g C); k_{deg} degradation rate of detritus in the bottom (d⁻¹); k_{den} denitrification rate (d⁻¹); k_m nitrification rate (d1); q - nutrient fraction which becomes when phytoplankter dies; a- the fraction of nutrient per unit of phytoplankton type; ko- correction factor for oxygen concentration; O- oxygen concentration in overlying water (g m⁻³); fro- fraction reduced substances permanently removed; NOc- critical nitrate concentration (g N m⁻³); F_{ber3}- burial flux based on bottom volume (m d⁻¹); Cac- adsorption capacity (g P kg-1 dry matter); bio- amplification factor for bio-irrigation (-)Fn, nitrate flux to the denitrifying layer (g N m-2d-1); PO-organic

2.3 Macrophyte model of Potamogeton pectinatus L

The macrophyte model describing Potamogeton pectinatus L. has been developed coupling with parameters taken from phytoplankton model. Schematization of shoot, secondary shoot, root, mother tuber and new tuber for Potamogeton pectinatus L is given in fig. 1.

phosphate concentration (g P m³); OC, oxygen consumption flux in the boundary layer (g O₂ m³d⁻¹); OC oxygen consumption rate (g O₂ m³d⁻ i); $\mathcal{P}_{p_{beam}}$ phosphate release flux from the boundary layer (g P m⁻²d⁻¹); $\mathcal{P}_{Q_{pre}}$ precipitated phosphate concentration (g P m⁻²d⁻¹); z_i the thickness of the i th layer (m) M the natural mortality of phytoplankton (per day); u the mineralization rate constant of detritus (per day); s the sedimentation rate constant (per day); d diffusion coefficient (m²d⁻¹); f adsorbed, dissolved, precipitated or organic phosphate fractions; dstoichiometric constant for nitrogen in detritus (g N/g C); & stoichiometric constant for oxidation of detritus (g O/g C); & stoichiometric constant for nitrification (g O/g N); kp phosphate precipitation rate (d⁻¹); Ks half saturation phosphate concentration (g P m⁻³ pore water)

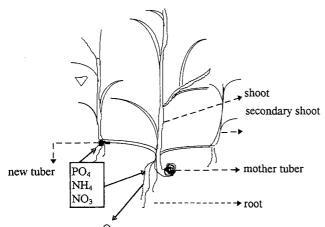


Fig. 1- The definition of shoot, secondary shoot, root, mother tuber and new tuber for Potamogeton pectinatus L.

The equations governing macrophyte development of root, shoot, secondary shoot, mother tuber and new tuber are presented in table 4.

Table 4. Model equations for macrophyte development in shallow lake

$$\frac{dB_{sht}}{dt} = Ph_{sht} - RESP_{sht} - (1 - C_{dead})DEAD_{sht} + TUBF - ELSEC. \max[0, F_{sht}] - f_{sht} \cdot e_{sht} \cdot B_{sht}$$

$$\frac{dB_{sec}}{dt} = Ph_{sec} - RESP_{sec} - (1 - C_{dead})DEAD_{sec} + ELSEC. \max[0, F_{sht}] + f_{sec} \cdot e_{sec} \cdot B_{sec}$$

$$\frac{dB_{sec}}{dt} = Ph_{sec} - RESP_{sec} - (1 - C_{dead})DEAD_{sec} + ELSEC. \max[0, F_{sht}] + f_{sec} \cdot e_{sec} \cdot B_{sec}$$

$$\frac{dB_{noot}}{dt} = RESP_{nob} - DECAY_{nob} - TUBF$$

$$\frac{dB_{noot}}{dt} - RESP_{nob} - DECAY_{nob} - TUBF$$

$$\frac{dB_{nob}}{dt} - GR_{nob} - RESP_{nob} - DECAY_{nob} - TUBF$$

$$\frac{dB_{nob}}{dt} - GR_{nob} - RESP_{soc} - (1 - C_{dead})DEAD_{soc} + RESP_{soc} - B_{sec}$$

$$\frac{dB_{nob}}{dt} - GR_{nob} - RESP_{nob} - DECAY_{nob} + NTUBF, f(f) + f_{sht} \cdot e_{sht} \cdot B_{sht} + f_{sec} \cdot e_{sec} \cdot B_{sec}$$

$$\frac{dB_{nob}}{dt} - GR_{nob} - RESP_{sht} - (1 - C_{dead})DEAD_{sht} + TUBF$$

$$\frac{dB_{nob}}{dt} - GR_{nob} - RESP_{sht} - (1 - C_{dead})DEAD_{sht} + TUBF$$

$$\frac{dB_{nob}}{dt} - RESP_{sht} - RESP_{sht} - (1 - C_{dead})DEAD_{sht} + TUBF$$

$$\frac{dB_{nob}}{dt} - RESP_{sht} - RESP_{sht} - (1 - C_{dead})DEAD_{sht} + TUBF$$

$$\frac{dB_{nob}}{dt} - RESP_{sht} - RESP_{sht} - (1 - C_{dead})DEAD_{sht} + TUBF$$

$$\frac{dB_{nob}}{dt} - RESP_{sht} - RESP_{sht} - (1 - C_{dead})DEAD_{sht} + TUBF$$

$$\frac{dB_{nob}}{dt} - RESP_{sht} - RESP_{sht} - (1 - C_{dead})DEAD_{sht} + TUBF$$

$$\frac{dB_{nob}}{dt} - RESP_{sht} - RESP_{sht} - (1 - C_{dead})DEAD_{sht} + TUBF$$

$$\frac{dB_{nob}}{dt} - RESP_{sht} - RESP_{sht} - (1 - C_{dead})DEAD_{sht} + TUBF$$

$$\frac{dB_{nob}}{dt} - RESP_{sht} - RESP_{sht} - (1 - C_{dead})DEAD_{sht} + TUBF$$

$$\frac{dB_{nob}}{dt} - RESP_{sht} - RESP_{sht} - (1 - C_{dead})DEAD_{sht} + TUBF$$

$$\frac{dB_{nob}}{dt} - RESP_{sht} - RESP_{sht} - (1 - C_{dead})DEAD_{sht} + TUBF$$

$$\frac{dB_{sec}}{dt} - RESP_{sht} - RESP_$$

where B_{sh} shoot biomass (g.m²); S_{sec} secondary shoot biomass (g.m²); S_{roo} root biomass (g.m²); B_{nb} tuber biomass (g.m²); $P_{h_{sht}}$ gross photosynthesis of shoot (g.m²d⁻¹); $P_{h_{sec}}$ gross photosynthesis of secondary shoot (g.m²d⁻¹); $RESP_{shc}$ respiration of shoot (g.m²d⁻¹); C_{shod} fraction of dead biomass available for growth; $DEAD_{shc}$ dead of shoot biomass per day (g.m²d⁻¹); $RESP_{sec}$ respiration of secondary shoot (g.m²d⁻¹); P_{shc} dead rate of secondary shoot (d⁻¹); P_{shc} biomass of tuber coming from tuber for shoot growth per day g.m²d⁻¹); P_{shc} dead rate of shoot (d⁻¹); P_{shc} respiration of tuber bank biomass available for biomass initiation (d⁻¹); P_{shc} respiration rate of shoot (d⁻¹); P_{shc} respiration of tuber bank (d⁻¹); P_{shc} dead rate of root (d⁻¹); P_{shc} respiration of tuber bank (d⁻¹); P_{shc} dead rate of root (d⁻¹); P_{shc} insolation in PAR (photosynthetically active radiation) halfway averaged over the day (µE m⁻²d⁻¹); P_{shc} insolation in PAR (photosynthetically active radiation) halfway averaged over the day (µE m⁻²d⁻¹); P_{shc} half saturation constant for nitrogen concentration in pore water sediment (g.N. m⁻³); P_{shc} half saturation of PAR; P_{shc} fraction of shoot biomass transfered to new tuber biomass; P_{shc} fraction of shoot biomass transfered to new tuber biomass; P_{shc} fraction of shoot biomass (g.m²d⁻¹); P_{shc} maximum dead rate of secondary shoot biomass (d⁻¹); P_{shc} fraction constant for new tuber biomass (d⁻¹); P_{shc} fraction constant for shoot biomass (d⁻¹); P_{shc} fraction of new tuber biomass (d⁻¹); P_{shc} fraction of shoot biomass (d⁻¹); P_{shc} fraction of new tuber biomass (d⁻¹); P_{shc} fraction of new tuber biomass (d⁻¹); P_{shc} fraction constant for shoot biomass (d⁻¹); P_{shc} half saturation constant for shoot biomass (d⁻¹); P_{shc} half saturation constant for shoot

3. Model verification and application

The verification was conducted using observed data of ammonium, nitrate, phosphorus, three species of algae (diatom, green algae and blue green algae), measured soluble reactive phosphorus concentrations, total biomass of macrophyte (Potamogeton pectinatus L) in Lake Veluwe (The Netherlands) from January to November, 1986 (figs 2-5). The Lake has the surface area of about 32.8 km² with average depth of 2m. The decline of Potamogeton pectinatus L. in the Lake are considerable from 1969 to 1975 (Schffer at al., 1994) due to the lake eutrophication. From 1979 the restoration measures were taken by reducing external phosphorus loading from 3.0 to 1.0 g P m² y¹ with retention time of dissolved

compounds from 0.5 to 0.25 years (van der Molen al al., 1994). As the results, diatoms and green algae were dominant from 1985 and Potamogeton pectinatus L. gradually recovered.

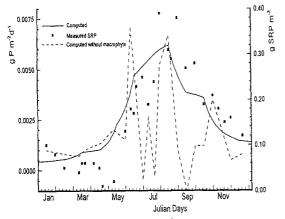


Fig. 2: The comparison between phosphate return fluxes with and without including macrophyte, measured SRP in Lake Veluwe in 1978.

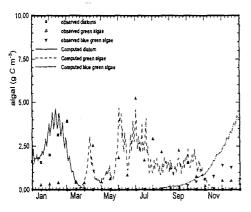


Fig. 4: The comparison between measure and computed concentration of diatoms, green and blue green algae in Lake Veluwe in 1986

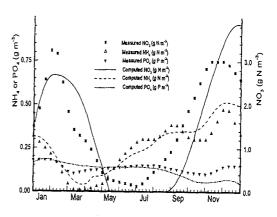


Fig. 3: The comparison between measured and computed values of nitrate, ammonium and phosphorus in Lake Veluwe in 1986.

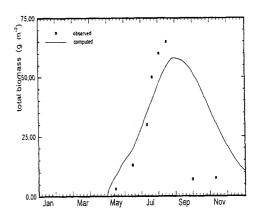


Fig. 5: The comparison between observed and computed of total macrophyte biomass in Lake Veluwe in 1986.

It is evident that the discrepancies between measured and calculated data are in the acceptable ranges. It is noted that phosphorus return flux calculated in the model gives better agreement with measured SRP than that calculated without including macrophyte (fig. 2).

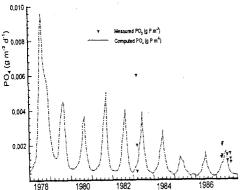


Fig. 6: The comparison between computed and measured phosphate return fluxes in Lake Veluwe from 1978 to 1987

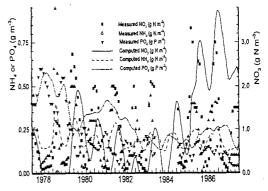
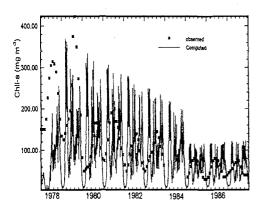


Fig. 7: The comparison between measured and computed values of nitrate, ammonium and total phosphorus in Lake Veluwe from 1978 to 1987

Keeping all the model parameters, long term investigation of nutrients, phytoplankton concentrations and macrophyte development in Lake Veluwe from 1978 to 1987 has reproduced (figs 6-9). Obviously by declining phytoplankton concentration from 1978 to 1987 (fig. 8), water clarity is improved, as a result, macrophyte was largely enhenced (fig. 9).



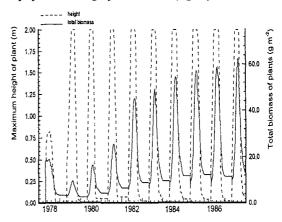


Fig. 8: The comparison between measured and computed concentrations of Chlorophyll-a in Lake Veluwe from 1978 to 1987

Fig. 9: Total macrophyte biomass and maximum height in Lake Veluwe from 1978 to 1987.

3. Conclusions

The model results show that the model was capable of reproducing real picture of phytoplankton and macrophyte development, nutrient dynamics in Lake Veluwe for short and long term period. By reducing external phosphate loading, phytoplankton compositions in Lake Veluwe were reduced, from nitrogen limited to phosphate limited and switched from blue- green algae to green algae and diatom. The decline of phytoplankton concentrations followed the water quality is improved, as consequences light can be penetrated into deeper water column and macrophyte can be grown. The model also can be used for analyzing the interactions of phytoplankton growth, macrophyte development and nutrient dynamics in other shallow lake for lake manager.

References

- 1 Carpenter, S.R., and Lodge, D.M., 1986, Effects of submersed macrophyte on ecosystem processes, Aquatic Botany, 341-370.
- 2 Smits, J.G.C., van der Molen, D.T., 1993, Application of SWITCH, a model for sediment water exchange of nutrients, to Lake Veluwe in The Netherlands, *Hydrobiologia*, **253**,281-300.
- 3 Hootsmans, M.J.M., 1994, Lake Veluwe, a Macrophyte-dominated System under Eutrophication Stress, *Kluwer Academic Publishers*, 250-286.
- 4 Scheffer, M. et al., 1994, Lake Veluwe, a Macrophyte-dominated System under Eutrophication Stress, Kluwer Academic Publishers, 20-25.
- 5 Mitchell, S.F., 1989, Primary production in a shallow eutrophic lake dominated alternately by phytoplankton and by submerged macrophytes, *Aqua. Bontany*, **33**, 101-110.
- 6 Hamilton, D.P. and Schladow, S.G., 1994a, Prediction of water quality in lakes and reservoirs: Part II: model description, Report No ED 597DH, Center for Water Reseach, University of Western Australia, 29 pp.
- 7 Chapell, A., 1995, A preliminary model of nutrient cycling in sediments of a Mediterranean lagoon, *Ecol. Modelling*, 80, 131-147.